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主要杂环胺类化合物研究进展

Research progress of the major HAAs

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摘要:对业界关于杂环胺的分类及其危害、烟气杂环胺的检测、主要杂环胺代谢产物和杂环胺暴露量监测的研究进展进行了综述,指出:杂环胺具有高致癌和致突变能力,烟气中杂环胺主要有10种,含量较高的为Harman, Norharman, A α C和MeA α C;对烟气中主要杂环胺的研究多集中在A α C和MeA α C,关于它们的代谢途径已经明确,但关于它们代谢产物数量的报道稍有差别;杂环胺暴露量监测主要通过检测分析人体尿液或毛发中的原型杂环胺实现,杂环胺代谢产物可作为暴露标志物用于杂环胺暴露量监测.未来的研究围绕Harman和Norharman的协同诱变作用、杂环胺A α C和MeA α C的代谢产物作为暴露标志物、快速有效地检测生物样本中杂环胺的方法和吸烟与杂环胺暴露的相关性等方面进一步开展,从而减少杂环胺的危害性,为“吸烟与健康”问题的研究提供参考.

关键词:

杂环胺;代谢产物;暴露量监测

Key words:

heterocyclic aromatic amines; metabolites; exposure level monitoring

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Abstract: The classification and hazard of HAAs, the detection of heterocyclic aromatic amines (HAAs) in cigarette smoke, the metabolism research progress of main HAAs and monitoring methods of human exposure were reviewed. It was pointed out that HAAs were highly carcinogenic and mutagenic. There are ten main HAAs in cigarette smoke, and the higher contents are Harman, Norharman, A α C and MeA α C. Studies of major heterocyclic amines in cigarette smoke are mostly focused on A α C and MeA α C. The metabolic pathways of A α C and MeA α C have been clarified, but there is a slight difference in the literature regarding the amount of their metabolites. The exposure monitoring of HAAs is mainly achieved through the detection of prototype HAAs in human urine or hair, metabolites of HAAs can be used as exposure markers for monitoring the exposure of HAAs. Future research will focus on the synergistic mutagenesis of Harman and Norharman, metabolites of A α C and MeA α C as exposure markers, rapid and effective detection method to determine HAAs in biological sample and the correlation between smoking and the exposure of HAAs to further reduce the harmfulness of HAAs and provide references for the study of "smoking and health".

0 引言

杂环胺 HAAs (heterocyclic aromatic amines) 是一类含有 N 杂环的多环芳香族化合物,由自由氨基酸、肌氨酸、肌酸酐与糖类高温反应产生,具有致癌、致突变活性^[1-2]。杂环胺广泛存在于煎炸食品^[3-6]、咖啡饮料^[6-7]、酒类^[8]、卷烟烟气^[9-12]、人体体液^[13-14]、河水和大 气^[15-16]中。自从 T. Sugimura 和其同事在 1977 年首次从食物中发现具有致癌活性的杂环胺以来^[5,17],目前已经有超过 25 种杂环胺类物质被分离鉴定出来^[1-2,4]。人体摄入杂环胺的渠道多种多样,但主要由食物摄入。卷烟烟气中也含有许多杂环胺,是人体摄入的杂环胺的重要来源,吸烟会导致某些杂环胺的暴露量上升,增加人们的健康风险。杂环胺被摄入人体或动物体之后,部分经 I 相和 II 相代谢形成解毒产物排出体外,也有一部分未经代谢直接以原型物的形态排出体外。评估杂环胺的暴露量主要是通过检测生物样本中杂环胺及其代谢物的含量来实现。本文拟综述杂环胺的分类与危害、烟气杂环胺的检测方法、主要杂环胺代谢产物和杂环胺暴露量的监测,以期“吸烟与健康”问题的研究提供参考。

1 杂环胺的分类与危害

1.1 杂环胺的分类

杂环胺主要产生于食物加热过程中。按照

形成过程,杂环胺可分为氨基咪唑杂环胺 AIAs (aminoimidazoazaarenes) 和氨基咔啉杂环胺 ACCs (amino-carboline congeners) 两类。AIAs 均含有咪唑环,为极性杂环胺,其 α 位置上有一个氨基,在体内经过代谢后形成 N-羟基化合物,具有致癌、致突变活性。这类化合物与 IQ 性质类似,环上的氨基均能耐受 2 mmol/L 的亚硝酸钠的重氮处理,所以该类化合物又称 IQ 型杂环胺,主要形成于 100 ~ 300 °C 温度环境,这类化合物主要是通过美拉德反应产生的,其主要杂环胺结构如图 1a) 所示。ACCs 又称热解杂环胺,为非极性杂环胺。该类杂环胺环上的氨基经 2 mmol/L 亚硝酸钠的重氮处理后脱落转变成 C-羟基化合物,所以又称为非 IQ 型杂环胺,它们的致癌、致突变活性较 IQ 型杂环胺弱。ACCs 形成温度较高,一般超过 300 °C^[2],主要来源于蛋白质和氨基酸的热裂解,其主要杂环胺结构如图 1b) 所示。

1.2 杂环胺的危害

杂环胺具有很高的致癌和致突变能力,动物实验研究表明^[1-2,4],它能诱导啮齿类和灵长类动物的多种器官产生肿瘤,并能引起哺乳动物的基因突变、染色体畸变和姐妹染色体互换。1993 年,国际癌症研究机构把 MeIQ, MeIQx, PhIP, A α C, MeA α C, Trp-P-1, Trp-P-2 和 Glu-P-1

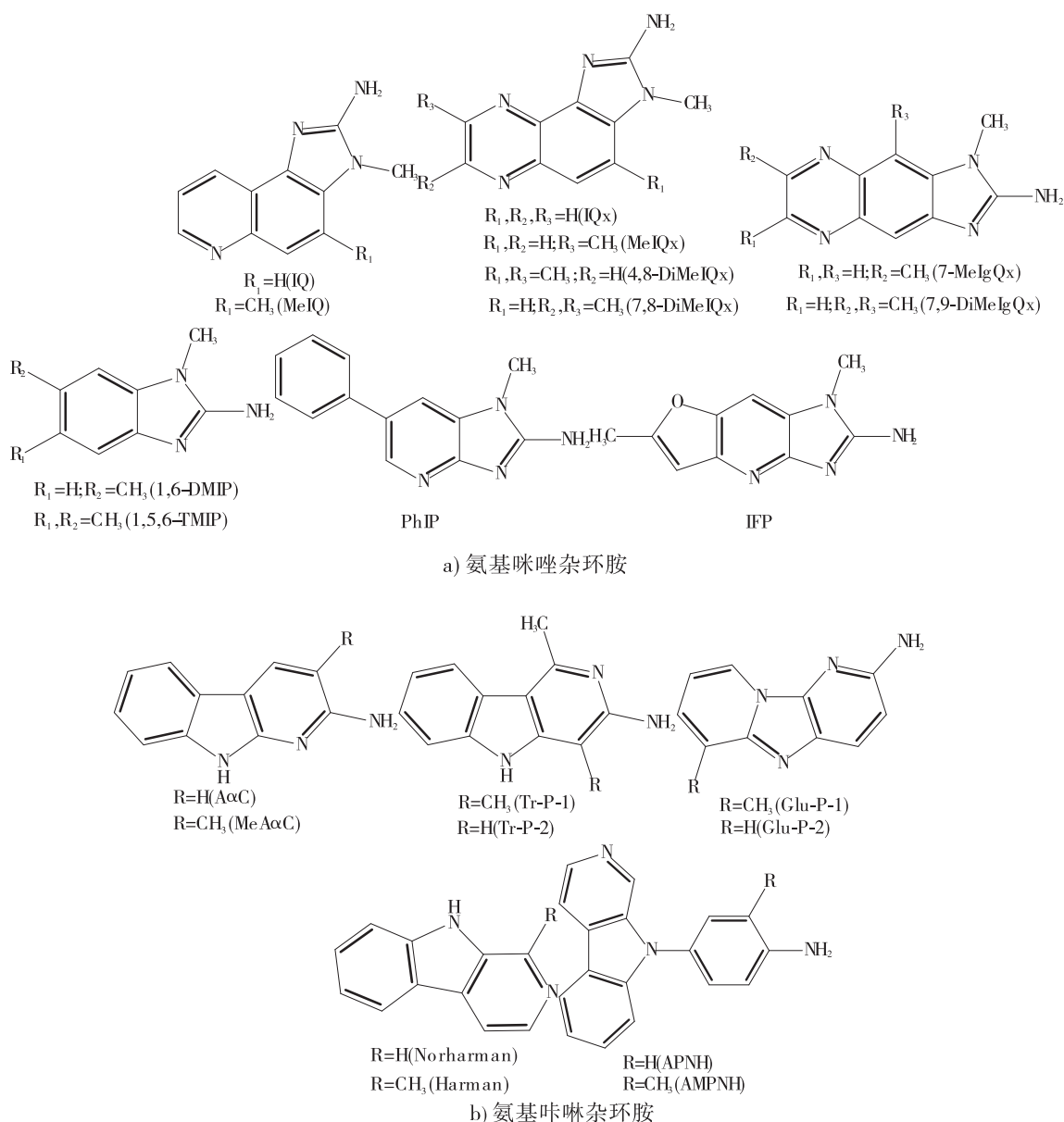


图1 常见杂环胺结构图

Fig.1 Chemical structures of prevalent HAAAs

这8种HAAAs作为2B类致癌物(潜在致癌物), IQ作为2A类致癌物(可疑致癌物)^[18];美国毒理学报告也把杂环胺类化合物列为人类的可能致癌物,建议减少其暴露量^[1-2]。

A α C和MeA α C是人体可疑致癌物.研究表明代谢活化的A α C和MeA α C对沙门氏细菌具有致突变活性^[11].G. Nauwelaers等^[19]用同样剂量的A α C,4-ABP,PhIP,MeIQx和IQ在人体肝细胞中培养,发现A α C与DNA的加合物的

量最多.Harman和Norharman为内生性化合物^[14,20-22],人体每天产生Harman和Norharman的量大约分别为每kg体重20ng和每kg体重50~100ng^[14].Harman和Norharman本身不具有致癌、致突变活性,但它们作为潜在的诱变剂或辅助致突变物会加强其他杂环胺的致癌、致突变能力^[23-26]。

2 卷烟烟气中杂环胺的检测方法

杂环胺类化合物是卷烟烟气中非常重要的

有害成分,主要是在卷烟抽吸过程中由含氮化合物和含氧化合物燃烧、裂解而产生的,在主流烟气中的释放量较高,单位通常为 ng/支. 对卷烟烟气中杂环胺的研究起始于1962年 Jr E. H. Poindexter 等^[27]报道的卷烟烟气中的 Harman 和 Norharman;1980年代, D. Yoshida 等^[28-29]报道卷烟烟气中含有 AaC 和 MeAαC. 1990年, Y. Kanai 等^[30]检测到卷烟烟气中含有 Glu-P-1 和 Glu-P-2;同年, S. Manabe 等^[31]检测到卷烟烟气冷凝物中含有 AαC, MeAαC, Trp-P-1 和 Trp-P-2. 1998年, D. Hoffmann 等^[32-33]将 AaC, MeAαC 等8种杂环胺类化合物列入卷烟烟气有害成分名单. 随着检测技术的发展,目前烟气中检测到的杂环胺主要有 Harman, Norharman, AαC, MeAαC, PhIP, IQ, Trp-P-1, Trp-P-2, Glu-P-1 和 Glu-P-2 这10种^[1,34-36].

卷烟烟气中杂环胺的测定主要采用剑桥滤片进行捕集^[37],再采用有机溶剂或酸性水溶液萃取滤片^[36],然后采用不同的前处理方法进行富集. 烟气杂环胺的富集纯化主要是用蓝棉^[35]、液液萃取^[36]、固相萃取^[38]和液液萃取串联固相萃取^[11,39]等方法,其中,固相萃取可以避免液液萃取过程中出现的乳化现象,提高萃取率,而且能有效地同时提取复杂样品中的多种杂环胺,是目前比较常用的前处理方法^[35,39]. 文献报道的分析方法主要依托 GC, GC-MS, LC 和 LC-MS 等检测手段;GC 和 GC-MS 需要将样品衍生化,操作较为繁琐,且检测限较低;烟气样品的复杂性,导致 LC 法样品选择性不高且不能同时检测多种化合物;而 LC-MS 以其高选择性和灵敏性越来越多地被应用于多种杂环胺的同时检测^[38-39]. H. Kataoka 等^[35]开发了一种 GC-NPD 检测烟气 AαC, Trp-P-1, IQ, PhIP 和 MeIQ 等6种杂环胺的方法,该方法采用液液萃取和蓝棉吸附再用 DMF-DMA 对杂环胺衍生,最后进行 GC 分析. T. A. Sasaki 等^[36]采用液液萃取、衍生化和 GC-MS-化学源负离子模

式检测卷烟主流烟气中的杂环胺 PhIP, IQ, AαC, MeAαC, Trp-P-1, Trp-P-2, Glu-P-1, Glu-P-2. 而 C. J. Smith 等^[12]应用固相萃取-GC-MS 方法分析了卷烟烟气中的 Harman, Norharman, AαC, MeAαC, 该方法无需衍生. G. Zhao 等^[38]采用固相萃取-HPLC-MS/MS 方法分析了卷烟烟气中 AαC, MeAαC, Trp-P-1 和 Trp-P-2 这4种杂环胺. 部分文献报道的烟气中10种杂环胺的释放量和检测方法见表1. 从表1中可以看出,卷烟烟气中的杂环胺以 Harman, Norharman, AαC 和 MeAαC 含量较高,除此以外,近年来常被报道的杂环胺还有 Trp-P-1 和 Trp-P-2,其余鲜见报道. 各篇文献所报道的卷烟烟气中杂环胺的释放量差异较大,这可能是由于受卷烟种类、抽吸方法、处理方法和检测手段等多种因素的影响造成的.

3 主要杂环胺代谢产物的定性分析

杂环胺有两条代谢途径解毒和活化^[1,4,17]: 第一步是被细胞色素 P450 酶催化的羟基化代谢过程,此为 I 相代谢, I 相代谢产物接着被酶催化,经历 II 相代谢,在这个代谢过程中,芳香环羟基化产物和部分环外氨基羟基化产物被葡萄糖醛酸转移酶 (UGTs) 和磺基转移酶 (SULTs) 催化,并与葡萄糖醛酸或磺酸等形成解毒的结合物,最后经尿液或粪便排出体外;另一部分环外氨基羟基化产物被乙酰转移酶 (NATs) 或磺基转移酶 (SULTs) 催化酯化形成活化产物,这些活化产物进一步异裂产生亲电子的芳基氮鎓离子中间物,其易与大分子物质 (如 DNA、蛋白质、多肽等) 形成加合物,产生致癌活性^[1,40].

杂环胺被摄入动物或人体之后,部分经 I 相和 II 相代谢形成解毒产物排出体外,也有一部分未经代谢直接以原型物排出体外^[41-44].

3.1 体内代谢产物

对于杂环胺的体内代谢研究,学者们多采

表1 部分文献报道的卷烟烟气中主要杂环胺的释放量和检测方法

Table 1 The content of heterocyclic aromatic amines in cigarette smoke and detection method in literatures

文献	前处理方法	检测方法	Harman	Norharman	A α C	MeA α C	PhIP	IQ	Trp-P-1	Trp-P-2	Glu-P-1	Glu-P-2
[30]	Liquid-liquid extraction	HPLC	—	—	—	—	—	—	—	—	0.37~0.89	0.25~0.88
[35]	blue cotton/rayon adsorption-liquid liquid extraction	GC-NPD	—	—	19.6~50.0	—	14.8	3.3~6.1	2.7~3.3	—	—	—
[36]	two-step derivatization	GC-NCI/MS	—	—	—	—	0.46~4.65	0.33~2.56	—	—	—	—
[12]	SPE	GC-MS	254~1025	675~2534	29.9~60.4	4.9~10.2	—	—	—	—	—	—
[9]	SPE	HPLC-MS/MS	—	—	25~260	2~37	—	—	0.29~0.48	0.82~1.1	—	—
[10]	solvent extracted	HPLC-MS/MS	630~1800	800~3300	33~95	2.0~6.1	—	—	0.3~1.9	1.2~4.6	—	—
[11]	Liquid-liquid extraction and SPE	GC-MS/MS	—	247~1736	21.2~25.3	1.1~6.36	—	2.1~7.2	—	1.0~4.0	—	1.1~2.0
[38]	SPE	HPLC-MS/MS	—	—	18.1~76.4	1.8~7.5	—	—	1.0~3.46	0.62~4.62	—	—

用肝微粒体作为模型,体内代谢产物主要为羟基化产物^[45-50]. 1982年, T. Niwa等^[45]研究了A α C在小鼠肝微粒体的代谢情况,采用HPLC-UV检测到了5种主要代谢产物,并经鉴定得出N₂羟基代谢产物具有致突变活性的结论. 1996年, H. Raza等^[47]研究了A α C在啮齿动物肝微粒体和人肝微粒体的代谢情况,采用HPLC-UV检测到了6种主要代谢产物,并采用核磁对代谢产物进行了鉴定,发现主要代谢产物为3-OH-A α C和6-OH-A α C. 1998年, H. Frandsen等^[48]采用HPLC-UV, HPLC-MS和核磁研究了MeA α C在PCB诱导过的小鼠肝微粒体的代谢情况,鉴定出MeA α C的3个主要解毒羟基代谢产物和1个具有致癌活性的活化产物. 2002年,他们^[49]通过HPLC-DAD结合质谱技术研究了A α C和MeA α C在人肝微粒体、PCB诱导过的小鼠肝微粒体和正常小鼠肝微粒体的代谢情况发现,A α C的主要代谢产物为3-OH-A α C和6-OH-A α C, MeA α C的主要代谢产物为6-OH-MeA α C, 3-CH₂OH-A α C和7-OH-MeA α C. 2008年, T. Herraiz等^[50]采用HPLC-UV和HPLC-MS研究了Harman和Norharman在细胞色素P450酶和肝微粒体的代谢动力学情况,发

现Harman和Norharman主要的代谢产物是6-OH产物和N₂氧化物,也有少量的3-OH产物.

近年来,出现了利用细胞模型研究杂环胺代谢产物的报道. 2007年, Z. X. Yuan等^[51]研究了A α C在雄性大白鼠动物模型、大白鼠肝细胞和人HepG₂肝细胞的细胞模型的代谢情况,共报道了17种代谢物,其中新发现了N-乙酰A α C, N-葡萄糖醛酸结合物和一个芳环羟基化产物. 2011年, G. Nauwelaers等^[52]研究了PhIP, MeIQx和IQ在人体肝细胞和小鼠肝细胞模型中其代谢产物(DNA加合物)的形成量,通过实验发现,在人体肝细胞中PhIP, MeIQx和IQ与DNA的加合物的量远远高于小鼠肝细胞,这表明用小鼠作为动物模型评价致癌物容易低估很多致癌物质的活性. 2012年, Y. Tang等^[52]利用人体肝细胞模型研究了UGTS酶催化A α C的代谢情况发现,A α C代谢成葡萄糖苷酸加合物. H. Frederiksen等^[53]还报道了用HPLC-MS对口服一定剂量MeA α C和A α C的大白鼠的肝脏、结肠、肾脏和心脏组织中MeA α C和A α C的代谢产物(DNA加合物)进行了定性和定量分析,发现MeA α C和A α C只与2-脱氮鸟苷形成DNA加合物,且肝脏中形成的DNA加合物最多.

3.2 体外代谢产物

现有文献对杂环胺体外代谢研究主要集中在 A α C 和 MeA α C, 研究方法主要是给动物喂食标样, 尔后对其排泄物进行检测. 2004 年, H. Frederiksen 等^[54] 采用 HPLC-MS 结合 UV 对喂食一定剂量 MeA α C 的小鼠 24 h 尿液和粪便进行了检测, 定性定量地分析了小鼠尿液和粪便中的 MeA α C 和其 11 种代谢物, 结果表明, 约有 21% 的 MeA α C 在粪便中, 34% 的 MeA α C 在尿液中, 尿液中未代谢的 MeA α C 约占 3.5%. 同年, 他们^[55] 又报道了采用同样的方法对喂食一定剂量的³H 标记的 A α C 小鼠的 24 h 尿液和粪便进行了检测, 定性定量地分析了 A α C 及其 7 种代谢物, 结果表明, 约有 12% 的 A α C 在粪便中, 32% 的 A α C 在尿液中, 未代谢的 A α C 约占 2.5%.

4 主要杂环胺类化合物暴露量监测

目前, 杂环胺暴露量监测主要是通过检测分析人体尿液或毛发中的原型杂环胺来实现, 杂环胺代谢产物的检测也日益受到关注.

4.1 尿液中原型杂环胺检测

从尿液中分离出杂环胺有许多技术手段^[1], 如溶剂萃取^[13,56]、固相萃取^[57]、极白棉处理尿液后经离子交换柱萃取^[58], 以及分子印迹技术^[43]和免疫亲和技术^[59].

1991 年, H. Ushiyama 等^[58] 报道了采用蓝棉和离子交换柱作前处理, 之后用 HPLC-FLD 检测人体 24 h 尿样中的 Trp-P-1, Trp-P-2, PhIP 和 MeIQx 这 4 种杂环胺, 发现它们的质量浓度范围为 0.03 ~ 1.97 ng/mL. 1995 年, H. Ushiyama 等^[60] 分析了人体尿液中 Harman 和 Norharman 的含量, 采用极白棉、阳离子交换柱等前处理方法, 用 HPLC-FLD 对尿液样品进行分析, Harman 和 Norharman 含量分别为 97.7 ~ 298 ng/mL 和 9.3 ~ 33.5 ng/mL. 1997 年, R.

Reistad 等^[61] 采用液液萃取串联蓝棉吸附前处理方法, 用 GC-MS 检测了尿液中的 PhIP, MeIQx 和 DiMeIQx, 发现样品酸解以后杂环胺的含量明显提高, 最高可以上升 32%. 1999 年, L. C. R. Kidd 等^[62] 采用免疫亲和萃取柱串联 HPLC-ESI-MS, 分析了不同人种尿液中 PhIP 含量的差别, 发现美国白人尿液中 PhIP 含量明显低于其他两个有色人种. 2004 年, S. Sentellas 等^[63] 采用液液萃取、硅藻土萃取和固相萃取的方法分离和富集了 15 种杂环胺类化合物, 采用毛细管电泳质谱进行检测, 但由于方法的检测限较高, 不能用于测定实际尿液样品. 2010 年, H-J. Cha 等^[64] 采用多重固相微萃取对尿液进行处理, 建立了尿液中的 IQ, MeIQ, MeIQx, PhIP, Glu-P-1, Glu-P-2, A α C 和 MeA α C 的 HPLC-MS/MS 检测方法. 2011 年, De F. Andres 等^[65] 建立了检测尿液中的非极性杂环胺化合物的固相微萃取-毛细管液相色谱荧光检测方法. 2014 年, Y. F. Fu 等^[66] 采用液液萃取串联固相萃取的前处理方法, 用 HPLC-ESI/MS/MS 分析了人体尿液中 15 种杂环胺的含量, 发现吸烟会导致 A α C 含量升高.

4.2 毛发中原型杂环胺检测

人体和动物的毛发也可以作为检测有害物质很好的基质, 毛发中杂环胺的含量能很好地反映杂环胺的暴露水平^[1,13,67-79]. 动物或人体摄入的杂环胺在组织、体液和排泄物中代谢时间较短, 而在毛发中代谢周期较长, 有文献报道在 4 周之后仍然能从毛发中检测到杂环胺的原型物^[1].

1999 年, R. Reistad 等^[70] 采用 GC-MS 检测人头发中的 PhIP, 检测到 12 个样本中 PhIP 的含量为 50 ~ 5000 pg/g. 2000 年, S. Hegstad 等^[71] 建立了一种固相萃取串联 GC-NCI/MS 检测人头发中 PhIP 的方法. 2005 年, M. Kobayashi 等^[72] 报道了一种柱切换的 LC-ESI/MS 在选择

离子扫描模式下检测人头发中 PhIP 的方法. 2009 年, E. E. Bessette 等^[73]报道了采用碱性水解、溶剂萃取和固相萃取分离人体毛发中 HAAs, 再采用 HPLC-ESI/MS/MS 进行检测的方法. 2013 年, H. Kataoka 等^[74]报道了一种在线固相微萃取串联 HPLC-ESI/MS/MS 检测人体头发中 16 种杂环胺含量的方法, 利用该方法对吸烟和非吸烟样本进行检测, 发现吸烟样本中的 IQ, MeIQx, Trp-P-1, PhIP 和 A α C 远高于非吸烟样本, 这 5 种杂环胺可以作为吸烟暴露生物标志物.

4.3 杂环胺代谢产物检测

关于尿液中杂环胺代谢产物的监测研究, 目前仅有 MeIQx 和 PhIP 代谢产物的报道. 2001 年, M. G. Knize 等^[75]报道了一种固相微萃取前处理结合 LC-MS/MS 分析技术检测人体尿液中 PhIP 的 4 种主要代谢物的方法, 分析发现 PhIP 的代谢受个体差异影响较大. 2002 年, H. Frandsena 等^[76]报道了用分子印迹固相萃取法结合 HPLC-UV 检测人体尿液中的 PhIP 代谢物 5-OH-PhIP; 同年, M. G. Knize 等^[77]报道了采用固相微萃取和 LC-MS/MS 技术分析人体尿液中 PhIP 及其 4 种主要代谢物的含量发现, PhIP 代谢有个体差异. 2008 年, H. Frandsen^[43]检测了人体尿液中 5-OH-PhIP, PhIP 和 4-OH-PhIP 的含量, 他们认为 5-OH-PhIP 能很好地反映 PhIP 的暴露量, 可作为其生物标志物. 2009 年, J. M. Fede 等^[78]给出了采用反向离子交换柱对尿液进行前处理, 然后用 HPLC-MS/MS 检测 PhIP 和其代谢物的方法. 2010 年, D. Gu 等^[79]给出了一种改进的 SPE 前处理富集人体尿液中的 MeIQx, PhIP 和它们的代谢物通过 HPLC-MS/MS 进行检测的方法.

对于杂环胺与 DNA 加合物的研究也有少量报道. 2010 年, E. E. Bessette 等^[80]应用 LC-ESI/MS/MSn 检测了人唾液中 PhIP, A α C 和

MeIQx 与 DNA 的加合物, 结果表明吸烟者唾液中 PhIP 与 DNA 的加合物检出比例较高. 2012 年, D. Gu 等^[81]采用 LC-ESI/MS/MSn 技术检测癌乳腺癌患者乳腺组织中的 PhIP 与 DNA 的加合物, 发现其含量水平为每 10^9 个核苷酸中有 3 个加合物.

5 结语与展望

本文对业界关于杂环胺的分类及其危害、烟气杂环胺的检测、主要杂环胺代谢产物和杂环胺暴露量监测等的研究进展进行了综述. 杂环胺作为强致癌和致突变物, 是卷烟烟气 Hoffmann 名单中重要有害成分, 其危害性越来越受到人们的重视.

目前, 文献报道的烟气杂环胺主要有 10 种, 杂环胺含量较高的为 Harman, Norharman, A α C 和 MeA α C 这 4 种, 这主要是因为卷烟燃烧锥的温度较高, 最高时甚至接近 $900\text{ }^{\circ}\text{C}$, 主要产生热解杂环胺即氨基咪唑杂环胺. 因此, 提高卷烟纸透气度、降低卷烟燃烧锥温度是降低烟气中杂环胺类化合物释放量的一条有效途径; 此外, 从杂环胺产生的根源出发, 降低烟叶中氨基酸、糖类和含氮化合物的含量, 也是降低烟气中杂环胺类化合物释放量的一条重要途径. 烹炸肉类食品是产生杂环胺类化合物的重要来源, PhIP, IQ 和 MeIQx 作为肉类食品中含量较高且致突变能力较强的化合物, 人们对这 3 种化合物的代谢研究较深入, 对它们代谢物的类型、代谢途径基本明确, 并已经寻找到相应的生物标志物, 而对于烟气中主要杂环胺的研究主要集中在 A α C 和 MeA α C, 现有文献对它们代谢产物的报道稍有差别, 还有待进一步研究. 作为诱变剂的 Harman 和 Norharman, 其协同诱变作用报道较少, 有待进一步研究.

杂环胺的强致癌活性使人们高度重视对其暴露量的研究, 目前杂环胺暴露量的监测主要

是检测分析人体尿液或毛发中的原型杂环胺,杂环胺代谢产物作为暴露标志物的研究也日益受到关注. 现有文献对杂环胺代谢产物作为暴露标志物的研究仅有关于 MeIQx 和 PhIP 代谢产物的报道,对于烟气中含量较高的 A α C 和 MeA α C 鲜见报道,还有待于进一步研究. 吸烟会导致部分杂环胺暴露量增加,随着“吸烟与健康”问题越来越多地受到公众的关注,吸烟与杂环胺暴露量的相关性正在被进一步地深入研究. 因此,探索快速有效地检测生物样本中杂环胺的方法,研究吸烟与杂环胺暴露量的相关性,将是今后的研究重点.

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